# PUBLICATION INFORMATION

This is the author's version of a work that was accepted for publication in the Nature Plants journal. Changes resulting from the publishing process, such as peer review, editing, corrections, structural formatting, and other quality control mechanisms may not be reflected in this document. Changes may have been made to this work since it was submitted for publication. A definitive version was subsequently published in https://doi.org/10.1038/s41477-018-0316-5

Digital reproduction on this site is provided to CIFOR staff and other researchers who visit this site for research consultation and scholarly purposes. Further distribution and/or any further use of the works from this site is strictly forbidden without the permission of the Nature Plants journal.

You may download, copy and distribute this manuscript for non-commercial purposes. Your license is limited by the following restrictions:

- The integrity of the work and identification of the author, copyright owner and publisher must be 1. preserved in any copy.
- 2. You must attribute this manuscript in the following format:

This is a pre-print version of an article by Hubau, W., De Mil, T., Van den Bulcke, J., Phillips, O.L., Angoboy Ilondea, B., Van Acker, J., Sullivan, M.J.P., Nsenga, L., Toirambe, B., Couralet, C., Banin, L.F., Begne, S.K., Baker, T.R., Bourland, N., Chezeaux, E., Clark, C.J., Collins, M., Comiskey, J.A., Cuni-Sanchez, A., Deklerck, V., Dierickx, S., Doucet, J.-L., Ewango, C.E.N., Feldpausch, T.R., Gilpin, M., Gonmadje, C., Hall, J.S., Harris, D.j., Hardy, O.J., Kamdem, M.-N.D., Kasongo Yakusu, E., Lopez-Gonzalez, G., Makana, J.-R., Malhi, Y., Mbayu, F.M., Moore, S., Mukinzi, J., Pickavance, G., Poulsen, J.R., Reitsma, J., Rousseau, M., Sonké, B., Sunderland, T.C.H., Taedoumg, H., Talbot, J., Tshibamba Mukendi, J., Umunay, P.M., Vleminckx, J., White, L.J.T., Zemagho, L., Lewis, S.L., Beeckman, H. 2019. The persistence of carbon in the African forest understory. Nature Plants, 5(2): 133-140.

DOI: https://doi.org/10.1038/s41477-018-0316-5



# 1 Title

## 2 The persistence of carbon in the African forest understory

3

## 4 Authors

5	Wannes Hubau <sup>1,2,3</sup> *, Tom De Mil <sup>1,2</sup> *, Jan Van den Bulcke <sup>2,4</sup> , Oliver L. Phillips <sup>3</sup> , Bhély Angoboy Ilondea
6	<sup>1,5,6</sup> , Joris Van Acker <sup>2,4</sup> , Martin J. P. Sullivan <sup>3</sup> , Laurent Nsenga <sup>1</sup> , Benjamin Toirambe <sup>1</sup> , Camille Couralet <sup>1</sup> ,
7	Lindsay F. Banin <sup>7</sup> , Serge K. Begne <sup>8,3</sup> , Timothy R. Baker <sup>3</sup> , Nils Bourland <sup>1,9,10,11</sup> , Eric Chezeaux <sup>12</sup> , Connie
8	J. Clark <sup>13</sup> , Murray Collins <sup>14</sup> , James A. Comiskey <sup>15,16</sup> , Aida Cuni-Sanchez <sup>17,18</sup> , Victor Deklerck <sup>2,4</sup> , Sofie
9	Dierickx <sup>1</sup> , Jean-Louis Doucet <sup>10</sup> , Corneille E. N. Ewango <sup>19,20,21</sup> , Ted R. Feldpausch <sup>22</sup> , Martin Gilpin <sup>3</sup> ,
10	Christelle Gonmadje <sup>23</sup> , Jefferson S. Hall <sup>24</sup> , David J. Harris <sup>25</sup> , Olivier J. Hardy <sup>26</sup> , Marie-Noel D. Kamdem
11	<sup>8,27</sup> , Emmanuel Kasongo Yakusu <sup>1,21,2</sup> , Gabriela Lopez-Gonzalez <sup>3</sup> , Jean-Remy Makana <sup>19</sup> , Yadvinder Malhi
12	<sup>28</sup> , Faustin M. Mbayu <sup>21</sup> , Sam Moore <sup>28</sup> , Jacques Mukinzi <sup>19,29</sup> , Georgia Pickavance <sup>3</sup> , John R. Poulsen <sup>13</sup> , Jan
13	Reitsma <sup>30</sup> , Mélissa Rousseau <sup>1,11</sup> , Bonaventure Sonké <sup>8</sup> , Terry Sunderland <sup>9,31</sup> , Hermann Taedoumg <sup>8</sup> , Joey
14	Talbot <sup>3</sup> , John Tshibamba Mukendi <sup>1,21,32</sup> , Peter M. Umunay <sup>33</sup> , Jason Vleminckx <sup>26,34</sup> , Lee J. T. White <sup>35,36,37</sup> ,
15	Lise Zemagho <sup>8</sup> , Simon L. Lewis <sup>3,17</sup> , Hans Beeckman <sup>1</sup>
16	
17	* These authors contributed equally to this work.
18	

## **19 Contact information**

- 20 <u>whubau@gmail.com</u>, <u>wannes.hubau@africamuseum.be</u>, <u>tom.demil@ugent.be</u>
- 21 Service of Wood Biology, Royal Museum for Central Africa
- 22 Tervuren, Belgium, Leuvensesteenweg 13, 3080 Tervuren
- 23

# 24 Affiliations

- 25 1 Service of Wood Biology, Royal Museum for Central Africa, Tervuren, Belgium
- 26 2 UGent-Woodlab, Laboratory of Wood Technology, Department of Environment, Ghent University,
- 27 Ghent, Belgium

- 28 3 School of Geography, University of Leeds, Leeds, UK
- 29 4 Centre for X-ray Tomography (UGCT), Ghent University, Ghent, Belgium
- Institut National pour l'Etude et la Recherche Agronomique, Kinshasa I, Democratic Republic of
   Congo
- 32 6 Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires
- 33 tropicaux (ERAIFT), Kinshasa, Democratic Republic of Congo
- 34 7 Centre for Ecology and Hydrology, Penicuik, UK
- 358Plant Systematic and Ecology Laboratory, Higher Teachers' Training College, University of
- 36 Yaounde I, Cameroon
- 37 9 CIFOR, Bogor, Indonesia
- 38 10 Forest Resources Management, Gembloux Agro-Bio Tech, University of Liege, Belgium
- 39 11 Resources and Synergies Development, Singapore, Singapore
- 40 12 Rougier-Gabon, Libreville, Gabon
- 41 13 Nicholas School of the Environment, Duke University, Durham, NC, USA
- 42 14 Grantham Research Institute on Climate Change and the Environment, London, UK
- 43 15 Inventory & Monitoring Program, National Park Service, Fredericksburg, VA, USA
- 44 16 Smithsonian Institution, Washington, DC, USA
- 45 17 Department of Geography, University College London, London, UK
- 46 18 Department of Geography and Environment, University of York, York, UK
- 47 19 Wildlife Conservation Society-DR Congo, Kinshasa I, Democratic Republic of Congo
- 48 20 Centre de Formation et de Recherche en Conservation Forestiere (CEFRECOF), Epulu, Democratic
   49 Republic of Congo
- 50 21 Faculté de Gestion de Ressources Naturelles Renouvelables, Université de Kisangani, Kisangani,
   51 Democratic Republic of Congo
- 52 22 Geography, College of Life and Environmental Sciences, University of Exeter, Exeter, UK
- 53 23 National Herbarium, Yaounde, Cameroon
- 54 24 ForestGEO, Smithsonian Tropical Research Institute, Panamá, Republic of Panama
- 55 25 Royal Botanic Garden Edinburgh, Edinburgh, UK

56	26	Service d'Évolution Biologique et écologie, Faculté des Sciences, Université Libre de Bruxelles,
57		Brussels, Belgium
58	27	Faculty of Science, Department of Botany and Plant Physiology, University of Buea, Buea,
59		Cameroon
60	28	Environmental Change Institute, School of Geography and the Environment, University of Oxford,
61		Oxford, UK
62	29	Salonga National Park, Kinshasa I, DR Congo
63	30	Bureau Waardenburg, The Netherlands
64	31	Faculty of Forestry, University of British Columbia, Vancouver, Canada
65	32	Faculté des Sciences Appliquées, Université de Mbujimayi, Mbujimayi, Democratic Republic of
66		Congo
67	33	Yale School of Forestry & Environmental Studies, New Haven, CT, USA
68	34	Department of Biological Sciences, Florida International University, FL, USA
69	35	Agence Nationale des Parcs Nationaux, Libreville, Gabon
70	36	Institut de Recherche en Ecologie Tropicale, Libreville, Gabon
71	37	School of Natural Sciences, University of Stirling, Stirling, UK
72		
73		
74		
75		
76		
77		
78		
79		
80		
81		
82		

83 Quantifying carbon dynamics in forests is critical for understanding their role in long-term climate regulation<sup>1,2,3,4</sup>. Yet little is known about tree longevity in tropical forests<sup>3,5,6,7,8</sup>, a factor that is vital for 84 estimating carbon persistence<sup>3,4</sup>. Here we calculate mean carbon age (the period that carbon is fixed in 85 trees<sup>7</sup>) in different strata of African tropical forests using (i) growth-ring records with a unique 86 87 timestamp accurately demarcating 66 years of growth in one site and (ii) measurements of diameter 88 increments from the African Tropical Rainforest Observation Network (23 sites). We find that in spite 89 of their much smaller size, in understory trees mean carbon age (74 years) is greater than in sub-90 canopy (54 years) and canopy (57 years) trees, and similar to carbon age in emergent trees (66 years). 91 The remarkable carbon longevity in the understory results from slow and aperiodic growth as an adaptation to limited resource availability<sup>9,10,11</sup>. Our analysis also reveals that while the understory 92 represents a small share (11%) of the carbon stock<sup>12,13</sup>, it contributes disproportionally to the forest 93 94 carbon sink (20%). We conclude that accounting for the diversity of carbon age and carbon sequestration among different forest strata is critical for effective conservation management<sup>14,15,16</sup>, and 95 for accurate modelling of carbon cycling<sup>4</sup>. 96

97

98 Investing in carbon storage and sequestration represent important climate change mitigation strategies<sup>3</sup>. Forests have a potential to provide both long-lived carbon stocks<sup>7,17</sup> and long-term carbon sinks<sup>1,2</sup>. To 99 successfully conserve forest carbon stock and increase forest carbon uptake, we must conserve carbon-rich 100 forests and extend the forested land area<sup>3</sup>, but decision makers and managers also need to understand the 101 long-term behaviour of carbon within forests<sup>1,2,3</sup>. Critical questions are: (i) how long does the carbon persist, 102 103 and (ii) where does it stay longest in the system? Carbon residence time is a direct function of tree longevity<sup>3,7,17</sup>, but attempts to estimate tree age in tropical rainforests are relatively scarce and often highly 104 contested<sup>5,6,7,8</sup>. Estimating the ages of the oldest trees in tropical forest stands is particularly subject to debate. 105 While some authors claim that broadleaved trees in the tropics may reach ages of 1000 years or older<sup>8,18</sup>, 106 others estimate maximum ages of not more than 600 years<sup>5,6</sup>. Furthermore, the oldest carbon in the system is 107 often assumed to be located in large trees<sup>8</sup>. Canopy and emergent trees contain a large proportion of the 108 stand-level biomass<sup>12,14</sup> but large trees alone may not represent well the entire forest in terms of growth rates, 109 tree lifespan and carbon persistence<sup>7</sup>. Canopy species grow faster<sup>19</sup>, but there is a general trade-off between 110

growth and lifespan in organisms<sup>9,20</sup>. Therefore, long-term carbon storage and sequestration in tropical
rainforests may substantially depend on smaller trees too.

113

Here, we take advantage of a remarkable rediscovery of a historic forest inventory plot to probe the age 114 115 structure of African rainforests in a way that has not been possible to date. The Nkulapark plot was 116 established in 1948 in the southwest of the Democratic Republic of the Congo (Supplementary Fig.1). A 117 total of 6315 trees with diameter on breast height (DBH) >=5cm were tagged and DBH was measured 118 annually for 9 years. In 2014, we rediscovered and measured 450 surviving tagged trees, of which 55 were 119 selected to measure growth ring series. We used the grown-in iron nail as a 1948 timestamp, giving accurate estimates of the DBH growth rate (in mm yr<sup>-1</sup>) and the rate of growth-ring formation (number of rings per 120 121 year) over a 66 year period for each tree (Fig.1, Supplementary Table 1). The age of each individual tree (in 122 years) was calculated as the total number of rings from pith to bark, divided by the rate of growth-ring 123 formation (number of rings per year) (Fig.1, Fig.2). We used the five-class Crown Illumination Index of Dawkins & Field (hereafter CII)<sup>21,22</sup> to compare growth patterns and tree age among the different rainforest 124 125 strata: understory (CII=1), sub-canopy (CII=2,3), canopy (CII=4) and emergent stratum (CII=5). Understory 126 trees receive no direct light, sub-canopy trees receive lateral light, canopy and emergent trees receive vertical 127 light (Fig.3a).Supplementary analyses show that the rediscovered Nkulapark tree dataset is adequate to 128 compare age differences between forest strata (Supplementary Fig.2, Supplementary Fig.3, Supplementary 129 Discussion).

130

We find that the age of the 55 Nkulapark trees with growth-ring series ranges between 129 and 452 years (Fig.2, Supplementary Table 1). There is no clear linear relationship between tree age and their DBH (p=0.082, Fig.2a). Understory trees (CII=1) do not differ significantly in age from canopy and emergent trees (CII=4 or 5) (p=0.254, boxplot at the right of Fig.2b), while trees in sub-canopy classes (CII=2 or 3) are slightly younger than trees in both the understory and the canopy. Despite their small size, understory trees (CII=1) can be surprisingly old. One Microdesmis puberula (TreeID=3545) has an estimated age of 329 years, with a DBH of just 156 mm (Supplementary Table 1).

139 To test if our findings hold true in a wider geographic context, we compared growth and age patterns among 140 the different forest strata in 23 Central African permanent forest inventory plots<sup>1</sup> (Fig.3, Table 1, Supplementary Fig.1 and Supplementary Table 2). Selected plots had a similar species composition as the 141 Nkulapark. Plots are demarcated rectangles or squares of median size 1 ha where each tree is mapped, tagged 142 and measured according to standard protocols<sup>1,2</sup>. DBH of each tree with DBH>= 10 cm was measured at 143 144 least twice. Small trees that grew larger than 10cm during the monitoring period were recorded as recruits. 145 Trees that died were recorded. We used repeated diameter measurements to estimate the growth rate of each individual tree. We estimated tree age by dividing the final diameter (mm) by the growth rate (mm yr<sup>-1</sup>), 146 assuming a constant growth rate over the lifetime of a tree<sup>7</sup>. We evaluated the robustness of this age 147 148 estimation method using the rediscovered Nkulapark trees as a reference (Supplementary Fig.4). 149 Uncertainties are due to relatively short plot monitoring periods (average 9 years), yielding negative or zero 150 growth rates for 9.7% of the trees. To avoid unrealistic tree age estimates, we replaced slow growth rates by a 'minimum allowed growth rate', defined as the x<sup>th</sup> percentile of the growth rate distribution within each CII 151 152 class. Sensitivity analysis showed that x=25 returned a realistic tree age distribution for our dataset (Supplementary Table 3, Supplementary Discussion). Further analysis suggests that x may be lower if 153 154 monitoring periods are longer. Finally, we estimated tree-level mean carbon age as the average age of each year ring, weighted by the carbon content of the ring<sup>7</sup>, with a year ring sequence deduced from the growth 155 156 rate.

157

158 The mean tree age for the 23 plots across the Congo Basin ranged between 131 and 284 years, with an 159 overall mean of 229 years (95% bootstrapped confidence intervals: 212-244 years) (Table 1). Mean tree age 160 in the understory (CII=1) is estimated to be 262 years, which is significantly older than the overall mean (p<0.001) and older than the mean age of the sub-canopy (CII=3, 187 years, p<0.001), the canopy (CII=4, 161 194 years, p<0.001) and emergent classes (CII=5, 221 years, p=0.021) (Table 1). Furthermore, mean carbon 162 age at the tree level is 65 years (95% CI: 61-70). Carbon stored in the understory trees (CII=1) is estimated to 163 be on average 74 years (69-79), which is significantly older than the overall mean (p<0.001) and older than 164 carbon stored in the sub-canopy (CII=3, 54 years, p<0.001) and the canopy (CII=4, 57 years, p=0.001). 165

166 However, the difference between mean carbon age in the understory (CII=1) versus the emergent class

167 (CII=5, 66 years) is not significant (p=0.086) (Fig.3b and Table 1).

168

For each forest stratum within each plot, we calculated above-ground biomass-carbon stock (AGC-stock, in 169 Mg C ha<sup>-1</sup>) and the AGC-sink (Mg C ha<sup>-1</sup> yr<sup>-1</sup>) using standard methods<sup>1,2</sup> (Table 1). The AGC-stock 170 171 represents the carbon reservoir in the system while the AGC-sink represents the net change, calculated as AGC-productivity (additions to the system from tree growth) minus AGC-mortality (losses)<sup>17</sup>. The mean 172 AGC-stock in the understory (CII=1) is 19 Mg C ha<sup>-1</sup>, which is 11% of the total plot-level AGC-stock (Table 173 1). The mean AGC-sink in the understory (CII=1) is 0.21 Mg C ha<sup>-1</sup> yr<sup>-1</sup> (Fig.3c), which represents 21% of 174 the plot-level AGC-sink. In contrast, the sub-canopy classes (CII=2 and 3) together contribute about 24% to 175 176 the AGC-stock, but nothing to the AGC-sink (-0.01 Mg C ha<sup>-1</sup> yr<sup>-1</sup>). The understory thus contributes 177 disproportionally to carbon sequestration, considering its relatively small share in the stock, and compared to 178 the limited contribution of the sub-canopy classes.

179

#### 180 **Discussion**

181

Results from both the Nkulapark dataset (Fig.2) and the 23 inventory plots (Fig.3, Table 1) show that 182 understory (CII=1) and emergent (CII=5) trees are on average older than sub-canopy trees (CII=3). This 183 184 pattern can be explained by differences in species composition<sup>15</sup>. Understory 'specialist species' maintain low growth rates for long periods, resulting in relatively small DBH at older age. Their adaptations allow 185 186 them to survive in the understory without the need to invest in rapid growth. The sub-canopy classes (CII=2 187 and 3) are populated with suppressed 'canopy specialists' that survived a recruitment stage but didn't reach 188 the canopy yet. These trees are relatively young (Fig.3) and they experience limited light conditions, 189 resulting in mortality rates equaling productivity rates (Table 1). Canopy and emergent trees (CII= 4 and 5 respectively) are canopy specialists that have been able to grow rapidly for long periods<sup>23</sup>. This is in line with 190 ref.<sup>19</sup> who show that growth rates in emergent trees are high and increase continuously. To test the 191 192 assumption that high tree and carbon age in the understory may be due to a difference in species 193 composition, we classified species as (i) understory specialists, (ii) non-specialists and (iii) canopy specialists 194 (Supplementary Discussion). This confirmed that understory specialists are on average smaller (p<0.001) but 195 older (p=0.003) than canopy specialists (Supplementary Table 4).

196

Furthermore, the Nkulapark dataset shows that there is a significant negative relationship between tree age 197 198 and growth-ring formation rate (Fig.2b). 91% of the trees in the dataset did not form a growth-ring every 199 year, suggesting an aperiodic growth pattern (shifts of growth to dormancy and back do not occur annually). 200 This aperiodic growth pattern is more prominent in understory trees, which formed fewer growth rings (0.36 201 rings per year) than canopy and emergent trees (0.61 rings per year) (p=0.01; boxplots at the top of Fig.2b). Aperiodic growth patterns translate into significantly slower long-term growth rates: understory trees (CII=1) 202 have a mean DBH growth rate of 1.52 mm yr<sup>-1</sup> versus 4.99 mm yr<sup>-1</sup> in emergent trees (p<0.001) (Table 1). 203 204 The observed differences in growth periodicity (Fig.2b), growth rates (Table 1) and age patterns (Fig.3) 205 among the different forest strata are most likely a result of differences in survival mechanisms which are a function of resource availability<sup>11,24, 25</sup> (Supplementary Discussion). 206

207

Our data show that growth and age distribution of tropical trees in mixed lowland rainforests is complex. 208 Large canopy trees are among the oldest in the rainforest, as suggested by several authors<sup>8</sup>, but as they 209 obtained their size and position by maintaining fast growth rates<sup>19</sup>, they are not significantly older than slow-210 growing small understory specialists (Fig.2, Fig.3, Table 1). Furthermore, while large canopy trees represent 211 the largest share of the carbon stock<sup>12,13,14</sup>, they suffer most during drought periods<sup>26</sup>. In comparison, 212 213 understory trees represent a smaller carbon capital but they are less vulnerable to drought and contribute 214 disproportionally to carbon sequestration. As such, the understory provides long-term stability in forest 215 carbon cycling. Furthermore, the understory is more diverse than the canopy in terms of species composition<sup>14,15,16</sup>. Therefore, we recommend quantifying forest ecosystem services by considering forests as 216 a whole, with all forest strata providing specific services<sup>16</sup>. This is important in Central Africa, where the 217 demand for fuelwood and charcoal could severely affect the understory if only large trees were to be 218 protected<sup>14,27</sup>. 219

220

221 Finally, our results suggest that care is required with large-scale modelling of forest carbon accumulation potential and responses to different climate change scenarios<sup>4</sup>. There are two hypothesised responses to 222 increasing atmospheric CO<sub>2</sub> concentrations, possibly explaining the long-term observed AGC-sink in tropical 223 forests<sup>1,2</sup>: (i) big trees increase their asymmetric competition benefit to the detriment of the rest of the stand, 224 or (ii) suppressed trees do best, as they are closer to their light compensation point<sup>17,28</sup>. Our results show that 225 226 both scenarios occur in forest stands, with the understory (CII=1), the canopy (CII=4) and the emergent (CII=5) classes contributing to carbon sequestration, while the sub-canopy classes (CII=2 and 3) do not 227 228 contribute. Forest and carbon cycle models will need to account for the diversity of carbon age and carbon sequestration potential among the different forest strata. Recent studies have found that forest structure can 229 be predicted from the characteristics of canopy trees<sup>13,14</sup>, but our results suggest that temporal dynamics 230 231 differ between forest strata. The long-term persistence of function depends on smaller trees too, which 232 compared to their stature contribute disproportionately heavily to long-term carbon storage, sequestration, 233 and climate resilience.

234

#### 235 **References**

- Lewis, S. L. et al. Increasing carbon storage in intact African tropical forests. Nature 457, 1003–1006
   (2009).
- 238 2. Brienen, R. J. W. et al. Long-term decline of the Amazon carbon sink. Nature **519**, 344–348 (2015).

239 3. Körner, C. A matter of tree longevity. Science **355**, 130–131 (2017).

4. Galbraith, D. et al. Residence times of woody biomass in tropical forests. Plant Ecol. Divers. 6, 139–
157 (2013).

5. Brienen, R. J. W., Schöngart, J. & Zuidema, P. A. Tree Rings in the Tropics : Insights into the
Ecology and Climate Sensitivity of Tropical trees. in Tropical Tree Physiology (eds. Goldstein, G. &

- 244 Santiago, L. S.) **6**, (2016).
- Worbes, M. One hundred years of tree-ring research in the tropics a brief history and an outlook to
  future challenges. Dendrochronologia 20, 217–231 (2002).
- Vieira, S. et al. Slow growth rates of Amazonian trees: Consequences for carbon cycling. Proc. Natl.
  Acad. Sci. 102, 18502–18507 (2005).

- 249 8. Chambers, J. Q., Higuchi, N. & Schimel, J. P. Ancient trees in amazonia. Nature **391**, 135–136
  250 (1998).
- 9. Bigler, C. Trade-Offs between growth rate, tree size and lifespan of mountain pine (Pinus Montana)
  in the swiss national park. PLoS One 11, 1–18 (2016).
- 10. Kleczewski, N. M., Herms, D. A. & Bonello, P. Effects of soil type, fertilization and drought on
  carbon allocation to root growth and partitioning between secondary metabolism and ectomycorrhizae
  of Betula papyrifera. Tree Physiol. **30**, 807–817 (2010).
- 11. Sass-Klaassen, U. Tree physiology: Tracking tree carbon gain. Nat. Plants 1, 15175 (2015).
- 12. Bastin, J.-F. et al. Seeing Central African forests through their largest trees. Sci. Rep. 5, 1–8 (2015).
- Bastin, J.-F. et al. Pan-tropical prediction of forest structure from the largest trees. Glob. Ecol.
  Biogeogr. in press, (2018).
- Lutz, J. A. et al. Global importance of large-diameter trees. Glob. Ecol. Biogeogr. 27, 849–864
  (2018).
- Memiaghe, H. R., Lutz, J. A., Korte, L., Alonso, A. & Kenfack, D. Ecological Importance of SmallDiameter Trees to the Structure, Diversity and Biomass of a Tropical Evergreen Forest at Rabi,
  Gabon. PLoS One 11, 1–15 (2016).
- Burton, J. I., Ares, A., Olson, D. H. & Puettmann, K. J. Management trade-off between aboveground
  carbon storage and understory plant species richness in temperate forests. Ecol. Appl. 23, 1297–1310
  (2013).
- Lloyd, J. & Farquhar, G. D. The CO2 dependence of photosynthesis, plant growth responses to
  elevated atmospheric CO2 concentrations and their interaction with soil nutrient status . I . General
  principles and forest ecosystems. Funct. Ecol. 10, 4–32 (1996).
- 18. Laurance, W. F. et al. Inferred longevity of Amazonian rainforest trees based on a long-term
  demographic study. For. Ecol. Manage. 190, 131–143 (2004).
- Stephenson, N. L. et al. Rate of tree carbon accumulation increases continuously with tree size.
  Nature 507, 90–93 (2014).
- 275 20. Wright, S. J. et al. Functional traits and the growth mortality trade-off in tropical trees. Ecology
  276 91, 3664–3674 (2013).

- 277 21. Synnott, T. J. A manual of permanent plot procedures for tropical rainforests. Tropical forestry
  278 papers 14, (1979).
- 279 22. Dawkins, H. C. & Field, D. R. B. A long-term surveillance system for British woodland vegetation.
  280 Oxford Forestry Institute Occasional Papers 1, (1978).
- 281 23. Hall, J. S., Harris, D. J., Medjibe, V. P. & Ashton, M. S. The effects of selective logging on forest
  282 structure and tree species composition in a Central African forest: Implications for management of
  283 conservation areas. For. Ecol. Manag. 183, 249–264 (2003).
- 284 24. Couralet, C., Van den Bulcke, J., Ngoma, L. M., Van Acker, J. & Beeckman, H. Phenology in
  285 functional groups of Central African trees. J. Trop. For. Sci. 25, 361–374 (2013).
- 286 25. Vico, G., Dralle, D., Feng, X., Thompson, S. & Manzoni, S. How competitive is drought
- deciduousness in tropical forests? A combined eco-hydrological and eco-evolutionary approach.
  Environ. Res. Lett. 12, 65006 (2017).
- 289 26. Bennett, A. C., McDowell, N. G., Allen, C. D. & Anderson-Teixeira, K. J. Larger trees suffer most
  290 during drought in forests worldwide. Nat. Plants 1, 15139 (2015).
- 27. FAO. *The charcoal transition : greening the charcoal value chain to mitigate climate change and*292 improve local livelihoods. (Food and Agriculture Organization of the United Nations, 2017).
- 28. Lewis, S. L., Malhi, Y. & Phillips, O. L. Fingerprinting the impacts of global change on tropical
  forests. Philos. Trans. R. Soc. B Biol. Sci. 359, 437–462 (2004).
- 295 29. Hietz, P. A simple program to measure and analyse tree rings using Excel, R and SigmaScan.
  296 Dendrochronologia 29, 245–250 (2011).
- 30. Benjamini, Y. & Hochberg, Y. Controlling the False Discovery Rate: A Practical and Powerful
  Approach to Multiple Testing. J. R. Stat. Soc. Ser. B 57, 289–300 (1995).
- 299

#### 300 Correspondence

- 301 Correspondence and requests for materials should be addressed to W.H. (<u>whubau@gmail.com</u>).
- 302
- 202
- 303
- 304

#### 305 Acknowledgments

306 W.H. and T.D.M. were both funded by the Brain program of the Belgian Federal Government

307 (BR/132/A1/AFRIFORD and BR/143/A3/HERBAXYLAREDD). The PhD project of T.D.M and the tenure 308 track of J.V.d.B. were supported by Ghent University Special Research Fund (BOF). Fieldwork was 309 sponsored by the King Leopold III fund for nature exploration and conservation, B.A.I. is supported by the 310 Institut National pour l'Etude et la Recherche Agronomiques en R.D.Congo (INERA- RDC- Luki) and the 311 Ecole Régionale Postuniversitaire d'Aménagement et de Gestion intégrés des Forêts et Territoires tropicaux 312 (ERAIFT Kinshasa). We thank WWF-RDC (Geert Lejeune), INERA and ERAIFT for facilitating fieldwork 313 in the Luki Reserve. We thank the INERA employees (Jean-Baptiste Ndunga, Jean-Maron, Fils Mbungu Phaka, Leonard Ngoma, Noble, Plaside), the WWF ecoguards and the students of the Universities of 314 315 Kinshasa (UNIKIN) and Boma for assistance in the field. For assistance with datasets we thank Marlène De Groot, Kévin Lievens, Piet Dekeyser, Stijn Willen and José Kempenaers. This paper is also a product of the 316 AfriTRON network, for which we are indebted to hundreds of institutions, field assistants and local 317 318 communities. This network has been supported by the European Research Council (291585, "T-FORCES"-319 Tropical Forests in the Changing Earth System, Advanced Grant to O.L.P.), the Gordon and Betty Moore Foundation, the David and Lucile Packard Foundation, the European Union's Seventh Framework 320 321 Programme (283080, 'GEOCARBON'), and Natural Environment Research Council (NERC) Consortium Grant 'TROBIT' (NE/D005590/1), 'BIO-RED' (NE/N012542/1) and a NERC New Investigators Grant, the 322 323 Royal Society, the Centre for International Forestry (CIFOR) and Gabon's National Parks Agency (ANPN). 324 We are indebted to the University of Yaounde I, the National Herbarium of Yaounde, Rougier-Gabon, the 325 Marien Ngouabi University of Brazzaville, WCS-Congo, Salonga National Park, WCS-D.R.Congo, and the 326 University of Kisangani for logistical support in Africa. 327

## 328 Author contributions

329 W.H., T.D.M., J.V.d.B., J.V.A. and H.B. conceived and designed the study. T.D.M. and B.A.I. coordinated

- 330 collection of Nkulapark data and wood cores. T.D.M. and J.V.d.B. measured growth ring series. W.H.
- 331 carried out the data analysis and wrote the paper. S.L.L., O.L.P., T.R.B. and Y.M. conceived the AfriTRON

332	forest census network programmes and the Forestplots.net database, and most co-authors helped collecting
333	AfriTRON forest census data. S.L.L., B.S., S.B., A.C.S., W.H., T.S. and L.W.W. coordinated forest plots
334	data collection. M.J.P.S., G.L.G., S.L.L., O.L.P., T.R.B. and G.P. contributed tools to analyze and curate
335	data. All co-authors commented on or approved the manuscript.
336	
337	Competing financial interests
338	The authors declare no competing financial interests.
339	
340	Figure legends
341	
342	Figure 1 Example of a wood core (Greenwayodendron suaveolens, TreeID=765) showing the 1948 nail
343	trace. The image at the top shows the full core. White lines indicate growth-ring boundaries, numbers
344	indicate growth-rings (counted from bark to pith), black arrows indicate important years. The bark to the
345	right of the figure indicates the year of sampling (2014). The dark discoloration in growth-rings 26 to 35 was
346	caused by oxidates from the iron nail that were transported up and down in damaged vessels and fibers. The
347	right border of the discoloration accurately marks the start of the year 1948. There are 25 rings between the

bark and the 1948 nail trace, suggesting that this individual needed on average 2.6 years to form a ring.

349 Using this rate for the 53 rings that were formed before 1948, we find that the first ring in the core was

probably formed around 1811. The location of the pith is indicated by the black lines to the left, which follow the direction of the wood rays<sup>29</sup>. This shows that the distance from the pith to the first ring boundary 351

352 in the core is about 11 mm. When using the average ring width of rings 78 to 68, we estimate that 7 rings are

353 missing. As such, this individual would be about 224 years old. The three close-ups at the bottom illustrate

354 wood anatomical details used to identify growth-ring boundaries (indicated by white triangles). Ring

355 boundaries in this species are demarcated by distended wood rays and flattening of the fibers. Black scale

357

356

bars cover 0.2 mm.

348

358 Figure 2 Variation in tree age inferred from growth-ring patterns in 55 trees where nail traces

accurately mark the year 1948. Panel (a) shows the relation between tree age and DBH. Panel (b) shows 359 the relation between tree age and the growth-ring formation rate (number of rings per year) between 1948-360 2014. In both panels, p-values of simple linear regression models are given in red. Brown dots represent 361 362 understory individuals (Crown Illumination Index =1), grey dots represent sub-canopy individuals (CII=2 and 3), blue dots represent canopy and emergent trees (CII=4 and 5). Boxplots show the first quartile, the 363 median value and the third quartile of the tree age distribution (vertical axis, boxplots to the right) and of the 364 365 variables in the x-axis (horizontal boxplots at the top of the panels). P-values under boxplots resulted from 366 two-sided Wilcoxon rank-sum tests. Outliers are marked with open circles.

367

Figure 3 Estimation of mean carbon age and AGC-sink per crown illumination category, for 23 368 369 permanent inventory plots in Central Africa. The 450 rediscovered Nkulapark trees were treated as an additional plot to estimate distributions of mean carbon age. Panel (a) illustrates the Crown Illumination 370 Index (CII) developed by Dawkins and Field (ref.<sup>22</sup>), with yellow arrows indicating reception of sunlight 371 (drawing modified from ref.<sup>21</sup>). Understory trees (brown, CII=1) receive no direct sunlight, sub-canopy trees 372 373 (grey) receive lateral (CII=2) or restricted vertical (CII=3) light, canopy trees (blue, CII=4) are almost 374 completely exposed to vertical light and emergent trees (blue, CII=5) have a crown that is completely exposed to vertical and lateral light. Boxplots in panels (b) and (c) show the distribution of plot-level mean 375 376 carbon age and plot-level AGC-sink per CII class. Boxplots represent the 25 % quartile, the median value and the 75 % quartile of the plot-level average ages. Outliers are marked with open circles. Comparison 377 378 among CII classes was performed with Dunn's rank-sum test using the Benjamini-Hochberg adjustment for multiple comparisons <sup>30</sup>. The grey dotted line shows the overall (plot-level) average age. 379

- 380
- 381
- 382
- 383
- 384
- 385

## 386 **Tables**

#### 387

metric	All trees	CII 1	CII 2	CII 3	CII 4	CII5	p-value
Tree age (yr)	<b>229</b> (212-244)	<b>262</b> (243-282)	<b>210</b> (196-223)	<b>187</b> (170-204)	<b>194</b> (175-212)	<b>221</b> (192-250)	0.021
Mean carbon age (yr)	<b>65</b> (61-70)	<b>74</b> (69-79)	<b>60</b> (55-64)	<b>54</b> (49-59)	<b>57</b> (51-63)	<b>66</b> (57-75)	0.086
AGC-prod (MgC ha-1 yr-1)	<b>3.1</b> (2.83-3.41)	<b>0.47</b> (0.4-0.56)	0.53 (0.42-0.67)	0.35 (0.27-0.43)	<b>0.9</b> (0.69-1.15)	<b>0.82</b> (0.63-1)	0.015
AGC-mort (MgC ha-1 yr-1)	<b>2.05</b> (1.66-2.48)	<b>0.26</b> (0.19-0.36)	<b>0.62</b> (0.49-0.76)	<b>0.28</b> (0.18-0.38)	<b>0.61</b> (0.41-0.85)	<b>0.27</b> (0.07-0.61)	0.058
AGC-sink (MgC ha-1 yr-1)	<b>1.05</b> (0.63-1.5)	<b>0.21</b> (0.09-0.33)	<b>-0.08</b> (-0.23-0.05)	<b>0.07</b> (-0.06-0.2)	<b>0.3</b> (0-0.59)	<b>0.55</b> (0.26-0.79)	0.011
AGC-stock (Mg C ha-1)	<b>177</b> (159-198)	<b>19</b> (15-23)	<b>26</b> (21-31)	<b>17</b> (12-22)	<b>58</b> (42-74)	<b>59</b> (46-75)	< 0.001
stem density (stems ha-1)	<b>428</b> (392-465)	<b>194</b> (165-228)	<b>112</b> (94-130)	<b>38</b> (33-44)	<b>59</b> (46-70)	<b>25</b> (19-31)	< 0.001
DBH (mm)	<b>313</b> (291-338)	<b>161</b> (154-168)	<b>235</b> (222-250)	<b>299</b> (273-326)	<b>435</b> (397-469)	<b>700</b> (644-760)	< 0.001
DBH growth (mm yr-1)	<b>2.38</b> (2.17-2.63)	<b>1.51</b> (1.35-1.69)	<b>2.16</b> (1.96-2.38)	<b>2.98</b> (2.69-3.31)	<b>4.03</b> (3.71-4.34)	<b>4.99</b> (4.32-5.68)	< 0.001
wood density (g cm-3)	<b>0.59</b> (0.56-0.61)	<b>0.64</b> (0.62-0.65)	<b>0.61</b> (0.6-0.63)	<b>0.59</b> (0.56-0.61)	<b>0.59</b> (0.56-0.61)	<b>0.57</b> (0.53-0.61)	0.006
ratio evergreen:deciduous	<b>3.27</b> (2.85-3.71)	4.72 (3.86-5.59)	<b>4.61</b> (2.76-7.85)	<b>3.62</b> (2.78-4.6)	<b>3.03</b> (2.09-4.22)	<b>1.78</b> (0.88-3.43)	< 0.001
proportion evergreen trees (%)	<b>55</b> (52-57)	<b>56</b> (52-59)	<b>55</b> (51-59)	<b>56</b> (52-61)	<b>51</b> (46-56)	44 (37-52)	0.01
proportion deciduous trees (%)	<b>18</b> (16-20)	<b>14</b> (12-17)	<b>19</b> (16-22)	<b>21</b> (17-25)	<b>26</b> (21-31)	<b>35</b> (28-41)	< 0.001

<sup>388</sup> 

#### 389 Table 1 Estimation of plot-level mean tree age, mean carbon age, aboveground carbon sink (AGC-

#### 390 sink), AGC-stock and leaf habit per crown illumination category, for 23 permanent forest plots. The

391 450 rediscovered Nkulapark trees were treated as an additional plot for estimation of tree age and mean

392 carbon age (first two rows). All metrics were averaged per plot and per CII class. Mean values are given in

393 bold, 95% confidence intervals are given between brackets. Components of the AGC-sink are AGC-

394 productivity and AGC-mortality. Components of AGC-stock are diameter (DBH), wood density and stem

395 density. Comparison among CII classes was performed with Dunn's rank-sum test using the Benjamini-

396 Hochberg adjustment for multiple comparisons<sup>30</sup>; the reported p-values compare CII1 and CII5.

- 397
- 398
- 399
- 400
- 401
- 402
- 403

#### 404 **Online Methods**

406 Site description. The Nkulapark is a phenology and tree-growth monitoring plot covering 174 ha within the 407 Luki Man and Biosphere reserve, located in the southern Mayumbe mountains in the Democratic Republic of the Congo<sup>24</sup> (Supplementary Fig.1). The region experiences a humid tropical climate with a dry season 408 409 between mid-May and mid-October and a short dry season from mid-December to mid-February. Yearly 410 precipitation ranges from 649 mm to 1853 mm with a mean precipitation of 1173 mm. Temperature ranges between 19 °C and 30 °C with a mean temperature of 25.5 °C <sup>24</sup>. The Nkulapark is situated almost entirely in 411 a catchment with a valley and a ridge and includes several microclimatic conditions. The semi-deciduous 412 413 lowland forest consists of (i) mature forest dominated by Prioria balsamifera, (ii) old regenerating forest dominated by Terminalia superba, (iii) mixed-species mature forest and (iv) modified forest patches<sup>31,32</sup>. 414 415 Nkulapark plot design in 1948. The Nkulapark was established and managed by the Institut National pour 416 417 l'Etude Agronomique du Congo Belge (INEAC), which was later renamed Institut National pour l'Etude et 418 la Recherche Agronomique en R.D.Congo (INERA, http://www.inera-drc.org). The person in charge of the tagging and the measurements was Léon Toussaint, who worked as a botanist in the Luki reserve between 419 1946 and 1952<sup>33</sup>. The planning of the plot was first announced in the INEAC-Luki annual report of 1946<sup>34</sup>. A 420 total of 29.2 km of observation paths were cut in the forest in 1947, following the contour lines of the Nkula 421 river valley (Supplementary Fig.1)<sup>31</sup>. A total number of 6315 trees were tagged by the end of  $1947^{31}$ , so we 422 assume that first wood formation after tagging occurred during the wet season that started in October 1948. 423 Tree selection was performed by randomly selecting trees from the pool of trees > 5 cm DBH in 1948. The 424 Nkulapark area was mapped, showing the locations of the largest tagged trees. From 1948 to 1957 yearly 425 diameter measurements were performed on all tagged trees<sup>35</sup>. Mortality events were recorded in the 426 datasheets. Trees were measured at a height of 1.3 m and the point of measurement (POM) was indicated on 427 428 the tree with a horizontal line of lead-based paint. For trees with buttresses or deformities, the POM was 429 raised to a point high enough to avoid the irregularities interfering with diameter measurements at 430 subsequent censuses. For trees with extremely high buttresses, diameters were estimated. For the same 6315 431 trees, weekly phenology observations were recorded. Phenological observations were done for leaf habit,

flowering, fruiting and seed dissemination. The plot was abandoned in 1957 but the datasheets were kept in
the library of the INERA station in Luki and digitized in 2008<sup>24</sup> and 2014.

434

Rediscovering Nkulapark trees in 2014. Each of the original 6315 Nkulapark trees were labelled during the 435 436 first census in 1948 with a zinc number tag that was attached to the tree using an iron nail of 8 cm long. A 437 part of these trees were indicated on the original 1948 map. During a first prospective field campaign in august 2014, this map was digitized and georeferenced with OGIS (OGIS development team, 2016) using 438 439 landmarks such as easily rediscovered trees, contour lines and observation paths that were still visible and 440 could be tracked with a Garmin 64s GPSmap (see Supplementary Fig.1). Based on this map, we pinpointed 441 the approximate location of 1521 individuals that were recorded as alive during the last census in 1957. 442 During a second field campaign in September- October 2014, these 1521 individuals were searched for. 450 443 of them were found alive, 16 were found dead and the remaining 1057 could not be relocated and were 444 assumed to be dead and rotten away, albeit some may have been missed (see Supplementary Fig.2, Supplementary Fig.3, Supplementary Discussion for an in-depth analysis of survivorship rates). The original 445 446 1948 tree tags and nails of the rediscovered trees were either still present outside the trunk or detected inside 447 the tree using a metal detector (BHJS, Bounty Hunter, USA). Scars on the trunk indicated the presence of an 448 overgrown nail, and repelled number tags were sometimes found on or in the ground nearby the tree using 449 the metal detector. In most cases, the numbers on the tags were still readable. On 95% of the rediscovered 450 trees, the lead-based paint of the POM was still visible, allowing a representative DBH measurement. 451 Comparison of DBH, DBH growth rates and tree age distribution in the original dataset of Nkulapark trees 452 versus the dataset of rediscovered trees, showed that the rediscovered dataset is slightly biased towards discovering slower-growing trees, but this bias affected both the classes of understory and canopy 'specialist' 453 species (Supplementary Fig.3). Hence, the rediscovered tree dataset is representative to compare growth and 454 455 age patterns among the different forest strata in the Nkulapark area.

456

457 Sampling Nkulapark trees in 2014. Wood samples for growth-ring analysis were taken from rediscovered
458 trees if following criteria were met : (a) the nail was still present in the wood, either totally grown-in or
459 partly sticking out of the trunk, (b) the exact position of the nail could be identified visually or with the metal

detector, (c) the nail was not overgrown by excessive wound tissue, buttresses or other deformities. As such, increment cores or stem discs were taken near the nail for 58 of the rediscovered trees. For each sampled tree, increment cores were taken a few centimetres above, below, to the left and to the right of the nail using a 40cm Pressler bore. For each tree, two additional cores were taken at 120° from the nail trace along the circumference of the tree. As such, 6 increment cores were available for each tree. This maximised the chance of sampling the pith of the tree. To study and describe the reaction of the wood after tagging, additional larger wood samples containing the nail were extracted from 30 trees using a saw and a chisel.

467

Visualizing and measuring growth-ring series. For each wood core, growth-ring series were visualized using 468 two imaging methods as described by ref.<sup>36</sup>: (i) first, density profiles were calculated from X-ray CT scans of 469 entire wood cores, then (ii) the cores were surfaced with a core microtome<sup>37</sup> and scanned using a flatbed 470 471 scanner (EPSON Perfection 4990 PHOTO). To obtain X-ray CT volumes, cores were scanned at 110 µm resolution with the Nanowood CT facility from the Centre for X-ray Computed Tomography of Ghent 472 University (UGCT, www.ugct.ugent.be)<sup>38</sup>, developed in collaboration with XRE (www.xre.be; now part of 473 the TESCAN ORSAY HOLDING a.s.). Reconstruction was performed with the Octopus software package 474 on a GeForce GPX 770 4GB GPU<sup>38,39</sup>. X-ray and flatbed scans were analysed using the toolchain for tree-475 ring analysis described by De Mil et al. (ref. <sup>36</sup>). This toolchain semi-automatically indicates the growth-ring 476 boundaries and calculates growth-ring width series. Depending on the visibility of the growth-ring patterns, 477 478 either the X-ray or the flatbed scans were used to check growth-ring boundaries and measure growth-ring 479 widths. Growth-ring boundaries were distinguished using visual wood anatomical characteristics such as distended rays, flattened fibers and terminal parenchyma bands<sup>5,40,41</sup>. For unclear ring boundaries, 480 481 microscopic thin sections were taken to study wood anatomy at high resolution using an Olympus BX60 482 microscope (Fig.1).

483

484 Detecting the 1948 nail trace. Discolorations or wound tissue formed as a reaction on the tagging were 485 visible in the cores taken near the nail. The surface of these stem discs was sanded to a few millimetres 486 above the nail and the anatomy was observed using an Olympus BX60 microscope. The nails were 487 remarkably well preserved inside the trees, probably due to cathodic protection of the iron by the zinc of the 488 tags. Hence, discolorations or wound tissue were visible in each of these samples. Discolorations were recognizable as darkened tissue in otherwise light-coloured wood (Fig.1). Discolorations occurred in cells 489 that were formed before the nail was inserted, due to oxidation processes between the wood and the iron nail. 490 Water in vessels and fibers in the neighbourhood of the nail (especially those damaged by the nail) spread 491 492 these oxidates upwards and downwards. Therefore, the discolorations in these cells are also detectable on 493 samples that were taken a few centimetres under or above the nails. Vessels and fibers that were formed after 494 the tagging were not damaged, hence did not show discolorations. As such, the boundary of the discoloration 495 accurately serves as a timestamp indicating the year of tagging (1948). Furthermore, wound-induced deformities occurred in the wood that was formed after the nail was inserted. This wound tissue is 496 497 characterized by increased woody productivity around the nail, forming a lump in the growth-rings that were 498 formed just after the nail was inserted. This lump formation is not present in the wood that was formed 499 before tagging.

500

501 Estimating tree age using growth rings and nail traces. Six cores were assessed for each tree. For some trees, 502 none of the cores contained the pith because (i) the tree radius exceeded the borer length or (ii) the pith was 503 eccentric and missed. In these cases, the core with the largest number of visible rings was used to estimate 504 the total number of growth-rings. The missing core length from the end of this core to the pith was estimated using the intersection of three lines of ring boundaries marked along the rays, as described by ref.<sup>29</sup> (see 505 506 Fig.1 for an example). We estimated the number of rings in the missing core part by dividing the missing 507 core length by the average ring width of the 5 oldest rings of the sampled core. We tested the robustness of 508 this method by rerunning the analysis using the average growth rates of the 5, 10, 15 and 20 oldest rings. We 509 found that the overall average tree age and the trends observed in Fig.2 are not sensitive to varying the 510 method used to estimate the number of rings in the missing part of the core. The core with the clearest nail 511 trace was used to count the number of rings formed after the year of tagging (1948). We used the number of growth-rings formed between 1948 (nail mark) and 2014 (cambium) as a reference to calculate the number 512 513 of years the individual needed to form one ring (years per ring). We then multiplied this ratio with the total 514 number of rings formed before 1948 to estimate the age of the tree (Fig.1 and Fig.2). Three individuals were

515 not used for analysis because the estimated missing core length of each core exceeded 20 cm. As such,

516 growth-ring series of 55 trees were retained (Supplementary Table 1).

517

Permanent forest inventory plots. To test if our findings hold true in a wider geographic context, we 518 519 estimated mean tree age and carbon age of the different forest strata in 23 Central African permanent forest 520 plots located in four different Central African countries (Cameroon, Gabon, Congo Brazzaville, D.R.Congo) (Supplementary Fig.1 and Supplementary Table 2). Plots were selected if at least 65% of the trees belonged 521 522 to species that also occur in the Nkulapark. Furthermore, plots selected for analysis conformed to the following criteria<sup>1,2,42</sup>: (1) plots had an actual plot area of  $\geq 0.2$  ha, (2) plots were georeferenced, (3) all trees 523 524 with DBH  $\geq$ 100 mm were measured, (4) the majority of stems were identified to species level, (5) plots had 525 at least 2 censuses, (6) plots had a total monitoring length of  $\geq$  3 years, (7) plots were situated within 526 structurally intact, apparently mature forest (excluding young or open forests), (8) plots were free from major 527 human impacts, (9) plots were located at  $\geq$ 50m from the anthropogenic forest edge, (10) altitude was below 1500 m.a.s.l., (11) mean annual air temperature was ≥20.0°C, (12) mean annual precipitation was ≥1000 mm 528 yr<sup>-1</sup>, (13) plots were located in terra firme forest. For analysis purposes, plots smaller than 0.5 ha that were 529 530 within 1 km of each other and located in similar forest types were merged (i.e. the LME cluster). All selected 531 plots are part of the African Tropical Rainforest Observatory Network (AfriTRON; www.afritron.org). These data are curated in the ForestPlots.net database<sup>43</sup>, and subject to identical quality control and quality 532 533 assurance procedures. All calculations of plot data metrics described hereafter were performed using the R statistical platform<sup>44</sup>, version 3.2.1. 534

535

Estimating AGC-stock. For each tree and each census, Aboveground Biomass at the tree level (AGB, Mg stem<sup>-1</sup>) was estimated using a published allometric equation for moist forests including terms for diameter (DBH, mm), dry wood density ( $\rho$ , g cm<sup>-3</sup>) and total tree height (H, m)<sup>45</sup> :

539 
$$AGB = \frac{0.0673 \times (\rho \times (\frac{DBH}{10})^2 \times H)^{0.976}}{1000}$$
 (formula 1).

540 Wood density values were derived from the dryad database (www.datadryad.org). Stems were matched to 541 species-specific wood density values or the mean values for the genus or family, following ref.<sup>1</sup> and<sup>43</sup>. 542 Heights were calculated using a single height-diameter model (Weibull) for central African lowland terra

543 firme forests published by ref.<sup>46</sup>, using commands implemented in the R-package BiomasaFP<sup>47</sup>.

544 Aboveground Biomass-Carbon (referred to as AGC) is considered as 47% of the AGB following IPCC

recommendations<sup>48</sup>. For each individual tree in the plot dataset, AGC-stock was calculated as the mean of the

546 first and last censuses.

547

548 Estimating AGC-sink. For the calculation of AGC-sink, only the first and the last censuses were used for each plot. First, AGC- productivity (Mg C stem<sup>-1</sup> yr<sup>-1</sup>) for each stem surviving the monitoring period, was 549 calculated as the difference between its total AGC at the end census minus the total AGC at the start census 550 551 of the interval, divided by the census interval length (yr). AGC-productivity for stems recruited during the 552 monitoring period (i.e. reaching DBH > 100 mm), was calculated in the same way, assuming DBH = 0 mm at the start of the interval. AGC-mortality for each tree that died during the monitoring period (Mg C tree<sup>-1</sup> yr<sup>-1</sup>) 553 was calculated as the AGC at the start of the monitoring period, divided by the total monitoring length (yr). 554 AGC-productivity at the stand level (Mg C  $ha^{-1}$  yr<sup>-1</sup>) was then calculated as the sum of tree-level productivity 555 estimates of all survivors and recruits. AGC-mortality at the stand level (Mg C ha<sup>-1</sup> yr<sup>-1</sup>) was calculated as the 556 557 sum of tree-level mortality estimates of all dead trees in the subset. We corrected for unobserved components of biomass growth and mortality due to census interval length effects, as discussed by ref.<sup>2</sup>, ref.<sup>49</sup> and ref.<sup>50</sup>. 558 A method to correct productivity and mortality rates for these uncertainties was developed by ref.<sup>49</sup> and 559 applied by ref.<sup>2</sup>. This correction accounts for (i) trees that recruit and die within the same interval (i.e. 560 561 unobserved recruits) and (ii) growth of trees that grow and die within the interval (i.e. unobserved biomass growth from mortality, which is not recorded because dead trees are not measured). As such, for each census 562 interval, we calculated unobserved recruitment and unobserved mortality components (Mg ha<sup>-1</sup> yr<sup>-1</sup>) using 563 the formulas proposed by ref.<sup>49</sup> and both components were added to both the AGC-productivity and AGC-564 565 mortality estimates. Estimates of the unobserved biomass components usually accounted for less than 3% of the total AGC-productivity and AGC-mortality. Finally, the AGC-sink was calculated as stand-level AGC-566 productivity minus AGC-mortality. Commands to calculate AGC-stock and AGC-sink are implemented in 567 the BiomasaFP R package <sup>47</sup>. 568

570 Tree age inferred from DBH growth rates in permanent inventory plots. For each tree within the permanent forest inventory plots, we estimated the age by dividing the DBH (mm) in the final census with the DBH 571 growth rate (mm yr<sup>-1</sup>) of the tree itself. For each tree, we averaged the DBH growth rate over all census 572 intervals preceding the last census. This method uses the actual growth rate of each tree, which is accurate 573 574 for healthy trees but returns unrealistic age estimates for trees with (i) slightly negative growth rates, (ii) zero growth rates or (iii) very slow growth rates. Such slow growth rates may be recorded in all DBH classes. 575 First, slow growth rates in large DBH classes may occur when a tree is diseased or at the end of its life 576 577 (senescence). These growth rates may not be representative for the total lifespan of these trees. Secondly, small growth rates may be recorded for small suppressed trees if their growth is so slow that it cannot be 578 579 recorded with sufficient precision using standard census procedures. Hence, these growth rates are replaced 580 by growth rates that are comparably small but yield a realistic tree age estimate. As such, we chose a 'minimum allowed growth rate' per CII class, following ref.<sup>7</sup>. We calculated the minimum allowed growth 581 rate for each CII class as the 25<sup>th</sup> percentile (=first quartile) of the growth rate distribution within the CII 582 class. For each tree with a growth rate slower than the minimum allowed rate, we replaced the growth rate by 583 the minimum allowed growth rate of the CII class. We conducted a sensitivity analysis to check how results 584 vary when varying the minimum allowed growth rate: we reran the analysis using the 10<sup>th</sup>, 15<sup>th</sup>, 20<sup>th</sup>, 25<sup>th</sup> and 585 30<sup>th</sup> percentile of the growth rate distribution within each CII class as a minimum allowed growth rate 586 (Supplementary Discussion and Supplementary Table 3). We used the average tree age in the dataset of 450 587 588 rediscovered Nkulapark trees as a reference to evaluate the tree-age estimation method based on DBH 589 growth rates (Supplementary Figure 4).

590

591 Mean carbon age. As a tree grows, it increasingly stores more carbon. Carbon near the bark of the tree is 592 younger than carbon in the pith. As such, mean carbon age of a tree does not equal total tree age. We 593 calculated mean tree-level carbon age using the same formula as ref.<sup>7</sup>:

594 mean carbon age = 
$$\frac{\sum_{i=1}^{n} (C_i \times A_i)}{\sum_{i=1}^{n} (C_i)}$$
 (formula 2),

595 with:

596  $C_i$  = carbon content of the i<sup>th</sup> ring (kg),

597  $A_i$  = age of the i<sup>th</sup> ring (yr),

598 n = nr of rings.

599

The carbon content of the i<sup>th</sup> ring is calculated as the carbon content of a tree with the DBH of ring i minus the carbon content of a tree with the DBH of ring i-1. To estimate carbon age in trees without measured growth-ring series (permanent inventory plots), we used the final DBH and the DBH growth rate (in mm yr<sup>-</sup> 1) to deduce a year-ring pattern, assuming that the growth rate was constant over the lifetime of the tree.

Classification of tree species and statistical analysis. We used the Nkulapark phenology data published by 605 ref.<sup>24</sup> to classify tree species as evergreen or deciduous. To distinguish between understory, sub-canopy, 606 607 canopy and emergent trees in the Nkulapark and the permanent forest plots, we used the Crown Illumination Index (CII) of Dawkins & Field (ref.<sup>22</sup>). Fig.3 illustrates the 5 classes with a drawing modified from ref.<sup>21</sup>. 608 The Crown Illumination Index was recorded in 23 of the selected permanent inventory plots. In each plot 609 610 where the index was recorded, each tree was attributed to one of the CII classes. The index was attributed in 611 the field, mostly during one census and mostly by a single person. We estimated mean tree age, mean carbon 612 age, mean AGC-stock and mean AGC-sink for each of the CII classes in each plot where the index was 613 recorded. None of the metrics reported in Table 1 meet the criterion of homogeneity of variances (Bartlett 614 test). Therefore, differences among the CII classes were tested using the non-parametric Dunn's rank-sum test. To avoid the multiple comparison problem, we used the Benjamini-Hochberg p-value adjustment<sup>30</sup> 615 ('dunn.test' package in  $R^{44}$ ). 616

617

Data availability. The input data and R-scripts to generate the figures and tables are available for download using the following private link : <u>https://figshare.com/s/06c793575d3b52ef5574</u>. Images of wood cores are available using the following link : <u>https://figshare.com/s/e6101fe7d330f8ea140a</u>. This folder also contains all annotation documents needed to visualize growth ring boundaries on the wood samples (please consult the README document for guidelines). Wood samples used to conduct this analysis are stored in the

623 Tervuren xylarium

624 (http://www.africamuseum.be/collections/browsecollections/naturalsciences/earth/xylarium). These samples

- 625 may be studied, within the Tervuren xylarium, upon request addressed to the curator H.B.
- 626 (hans.beeckman@africamuseum.be) or the corresponding author W.H. (whubau@gmail.com).

627

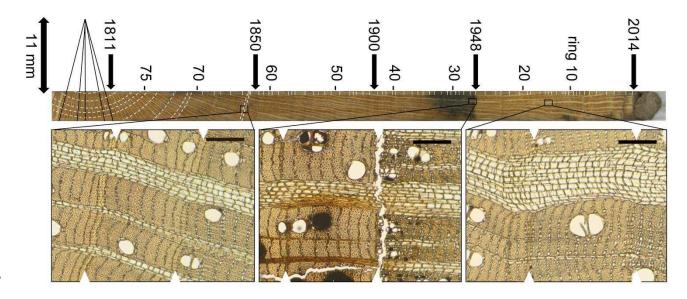
## 628 **References (methods only)**

- 629 31. INEAC. Rapport Annuel INEAC-Luki. (1947).
- 630 32. Coppieters, G. Inventaris van het archief van de Rijksplantages en de Regie der Plantages van de
- 631 Kolonie, het Nationaal Instituut voor de Landbouwkunde in Belgisch-Congo en de
- 632 Documentatiedienst voor Tropische Landbouwkunde en Plattelandsontwikkeling 1901-1999. (2013).
- 633 33. Académie Royale des Sciences d'outre-mer. Biographie Coloniale Belge/Biographie Belge d'Outre-
- 634 Mer, Tome IX. (2015).
- 635 34. INEAC. Rapport Annuel INEAC-Luki. (1946).
- 636 35. INEAC. Rapport Annuel INEAC-Luki. (1948).
- 637 36. De Mil, T., Vannoppen, A., Beeckman, H., Van Acker, J. & Van Den Bulcke, J. A field-to-desktop
  638 toolchain for X-ray CT densitometry enables tree ring analysis. Ann. Bot. 117, 1187–1196 (2016).
- 639 37. Gärtner, H. & Nievergelt, D. The core-microtome: A new tool for surface preparation on cores and
  640 time series analysis of varying cell parameters. Dendrochronologia 28, 85–92 (2010).
- 641 38. Dierick, M. et al. Recent micro-CT scanner developments at UGCT. Nucl. Instruments Methods
  642 Phys. Res. Sect. B Beam Interact. with Mater. Atoms **324**, 35–40 (2014).
- 643 39. Vlassenbroeck, J. et al. Software tools for quantification of X-ray microtomography at the UGCT.
- Nucl. Instruments Methods Phys. Res. Sect. A Accel. Spectrometers, Detect. Assoc. Equip. 580, 442–
  445 (2007).
- 646 40. Worbes, M. Tree-Ring Analysis. Encycl. For. Sci. 586–599 (2004).
- 41. Tarelkin, Y. et al. Growth-ring distinctness and boundary anatomy variability in tropical trees. IAWA
  J. 37, 275–294 (2016).
- 42. Phillips, O. & Baker, T. Field manual for plot establishment and remeasurement RAINFOR. Rainfor
  22 (2009). doi:10.13140/RG.2.1.1735.7202
- 43. Lopez-Gonzalez, G., Lewis, S. L., Burkitt, M. & Phillips, O. L. ForestPlots.net: A web application
- and research tool to manage and analyse tropical forest plot data. J. Veg. Sci. 22, 610–613 (2011).

- 44. R Development Core Team. R: A language and environment for statistical computing. (2008).
- 654 45. Chave, J. et al. Improved allometric models to estimate the aboveground biomass of tropical trees.
  655 Glob. Chang. Biol. 20, 3177–3190 (2014).
- 46. Feldpausch, T. R. et al. Tree height integrated into pantropical forest biomass estimates.

657 Biogeosciences **9**, 3381–3403 (2012).

- 47. Lopez-Gonzalez, G., Sullivan, M. & Baker, T. BiomasaFP package. Tools for analysing data
- downloaded from ForestPlots.net. R package version 1.1. Available at
- 660 http://www.forestplots.net/en/resources. (2015).
- 48. Aalde, H. et al. IPCC Guidelines for National Greenhouse Gas Inventories. Volume 4: Agriculture,
  Forestry and Other Land Use. Chapter 4: Forest Land. Forestry 4, 1–29 (2006).
- 49. Talbot, J. et al. Methods to estimate aboveground wood productivity from long-term forest inventory
  plots. For. Ecol. Manage. 320, 30–38 (2014).
- 665 50. Clark, D. a et al. Measuring net primary production in forest concepts and field methods. Ecol. Appl.
  666 11, 356–370 (2001).



667

